

Spill of Methyl Parathion in the Mediterranean Sea: A Case Study at Port-Said, Egypt

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Massive production and use of organic pesticides lead to the wide distribution of such toxic chemicals in the biotic and abiotic environment. Residues of pesticides in the aquatic ecosystem have raised much concern due to their hazardous effects on life and man.

Accidental spills of pesticides during their production and transportation may result in more serious and localized toxic effects (Lussen and Sechlimme 1971; Willmann 1976; Lafornera 1978).

In Egypt, the collision of two ships, namely "Garnet" and "Molaventure" took place near the northern entrance of Suez Canal at Port-Said, in February 1982. As a result, Garnet which was loaded with 31000 kg of methyl parathion slowly sank and was towed to shallow water. More than 10000 kg of methyl parathion found their way into the Mediterranean sea. Methyl parathion (O,O dimethyl-O-p-nitrophenyl phosphorothioate) is one of the most toxic organophosphorus insecticides (Webbe 1961; Morgan 1977).

The present work followed the concentration of methyl parathion in sea water, sediments and some fish species collected from the polluted area.

MATERIALS AND METHODS

The sampling area extended along a distance of approximately 5 km from the ship. Garnet was towed to an area defined by latitude $31^{\circ} 20'$ and longitude $32^{\circ} 20'$. Sub-surface water samples were collected from all sampling sites during the period of study (February 2nd to June, 14, 1982). Bottom sediments were obtained from the first three sampling sites. Composite fish and water samples were collected together from site No 4 along a distance of 2 km on both sides of the latter site, Fig 1.

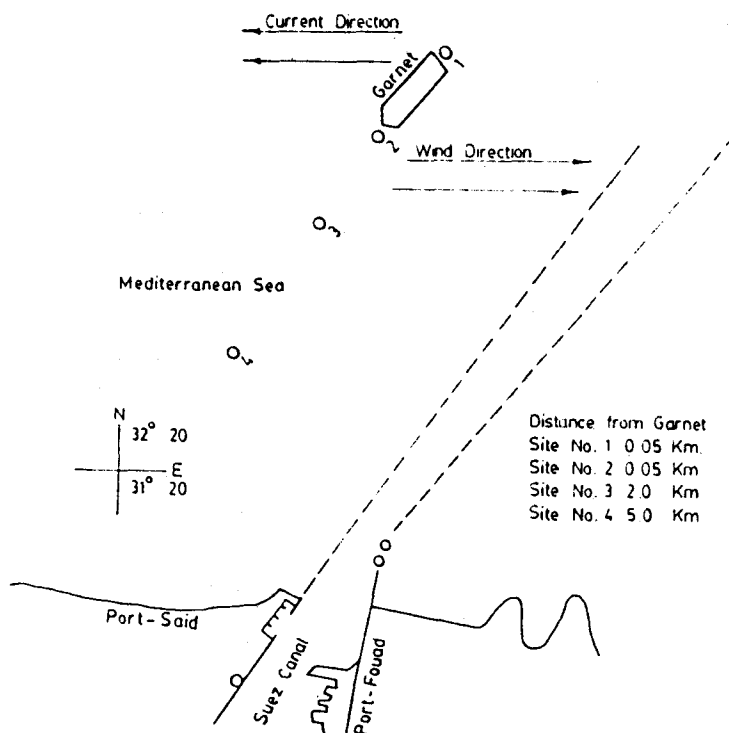


Fig 1. Schematic map showing sampling sites and location of Garnet

An adequate volume of sea water (250 mL) was transferred to a 500 mL separatory funnel and extracted twice with a 50 mL portions of 15% methylene chloride in hexane (v/v). The combined extracts were dried over anhydrous sodium sulphate and concentrated to 2 mL in a Kuderna-Danish evaporator.

A Known weight of air-dried sediments (50 g) was mixed with 50 g anhydrous sodium sulphate. The mixture was transferred to a glass chromatographic column. Further extraction and clean-up procedures were followed according to the US Environmental Protection Agency (EPA) method (1974).

The internal organs, head and tail of fish samples were removed and muscles were used for subsequent analyses. Extraction and clean up procedures were followed according to EPA method (1974).

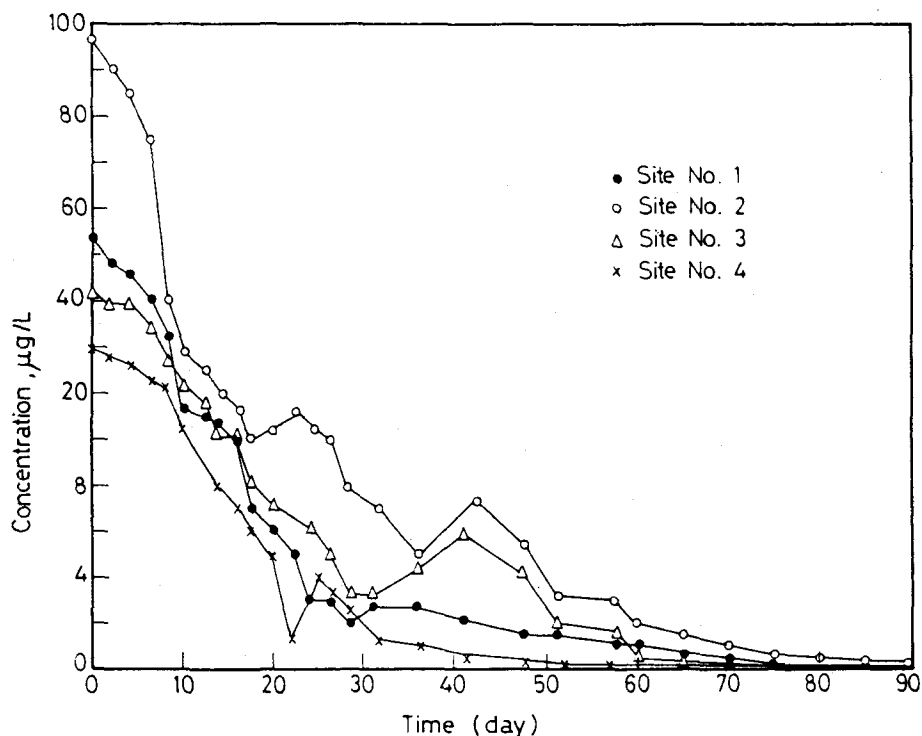


Fig 2. Concentration levels of methyl parathion in sea water at the various sites.

Residues of methyl parathion were identified and quantified using a Varian 3700 fitted with Ni^{63} electron capture detector and a glass column (4 mm I.D. and 2 meter length), packed with 3% OV 225 on 80/100 chromosorb W. The column, injector and detector temperatures were 180°C , 220°C , and 250°C , respectively. Nitrogen was used as a carrier gas at a flow rate of 40 mL/min. The detection limit, estimated as twice the signal to noise ratio, was $0.1\text{ }\mu\text{g/L}$.

RESULTS AND DISCUSSION

Results presented in Fig 2, indicate the general trend of distribution and concentration levels of methyl parathion in the studied area. Maximum concentration of $96\text{ }\mu\text{g/L}$ was reached at site No. 2. In general, the concentration of the insecticide decreased by distance and time to reach its lowest level of $\mu\text{g/L}$ after 50 and 80 days at sites No. 4 and 2, respectively. The follow up of the distribution of methyl parathion concentration at the various sampling

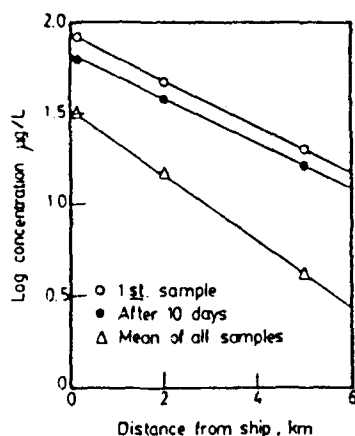


Fig 3. Logarithm of concentration of methyl parathion Vs distance.

sites indicated that the prevailing meteorological conditions (surface current and wind) have an influence in drifting the spilled insecticide towards the south - west direction.

In this study no attempt was made to measure the diffusion rate of the spilt insecticide. Several factors are also expected to take part in the decrease of methyl parathion concentration such as evaporation, chemical and biological degradation and adsorption on suspended matter and sediments (Kaufman and Plimmer 1972). A direct approach to assess the effects of dilution and transportation of pollutants, is the use of the first order decay equation: $C = C_i e^{-kt}$ where C_i = initial concentration C = concentration at time t and k is a constant (Baumgartner and Callaway 1972). On such basis,

Hashimoto et al (1978) developed the equation: $Y_o = Y e^{-kx}$ where Y

the concentration of pollutant at a distance "x" km; Y_o is the concentration at a source point, and "k" is a constant related to dilution. Such an equation leads to the derivation of the distance required to reduce the initial concentration of pollutant to its half value ($x_{\frac{1}{2}}$), where: $x_{\frac{1}{2}} = 0.693/k$. Results presented in Fig 3

indicate that obtainable data were in agreement with the given equation. The distance required to decrease methyl parathion concentration to its half value at site No.2, under the prevailing meteorological conditions, was 1.2 km during the first 10 days of sampling. However, when the mean values of concentrations attained at each sampling site were considered, the value of $x_{\frac{1}{2}}$ was only 0.8 km. Results

reached by this study add to the validity of the mathematical derivations given by Hashimoto et al (1978) in estimating the distance necessary to reduce the initial concentration of spilt insecticide to its half value.

Table 1 shows the concentration levels of methyl parathion in sediments together with its concentration in water collected simultaneously from the same sampling sites. It is evident that the residue levels in sediments gradually increased during the first 20 days of sampling. Residue levels of methyl parathion in sediments reached their highest values at site No. 2 whereas lowest values were found at site No 1. The decrease in residue levels in water coincide with the general increase in methyl parathion in sediments. Such a trend may be attributed to the uptake and adsorption of the insecticide by the suspended matter and sediments. The concentration factor given in Table 1 reached its maximum value of 49.5 at sampling site No.3, after 20 days of the spill. Meanwhile, values of the concentration factor at site No 3 were higher than its values at site No 2 which were still higher than values obtained at site No 1.

Residue levels of methyl parathion in water (Fig 2) and in sediments (Table 1) tend to show that the spilt insecticide was drifted towards the south - west direction under the influence of surface current and wind.

Variations in concentration levels of methyl parathion in fish and composite water samples collected from site No. 4 are presented in Table 2. Maximum concentration was found in the first sample of *Anguilla* species and amounted to 195.5 $\mu\text{g/L}$. Thereafter, residue of methyl parathion in *Anguilla* species decreased to 72.5 $\mu\text{g/kg}$. In case of *Sardine*, *Scidena* and *Mugil* species, residues of the insecticide attained their highest values in the second fish sample. Worth to be noted that the concentration of methyl parathion in sea water decreased by time whereas values of the concentration factor showed a general trend to increase and a maximum level of 6.9 was found in case of *Anguilla* species. In general, values of the concentration factor followed to order, *Anguilla* > *Mugil* > *Sardin* > *Scidena*.

Available results revealed that the uptake of methyl parathion was subject to variation according to fish species concerned, time of exposure and the concentration of the insecticide in water. According to Kanazawa (1978) the biococentration of diazinon in fish increased as the concentration of the compound and exposure time increased. Residue levels were also reported to decrease as the fish was transferred to clean water. Mathiessen et al (1982) indicated that residue levels of endosulfan in fish increased with increase of lipid content. Other investigators (Hamelink et al 1971

Table 1. Distribution of methyl parathion in sediments and sea water

Date of sampling	Concentration							
	Site No.1			Site No.2			Site No.3	
	Sediments µg / kg	Water µg/L	CF *	Sediments µg / kg	Water µg/L	CF	Sediments µg / kg	Water µg/L
18/2/82.	269.0	54.0	5.5	400.5	96.0	4.2	303.0	43.0
28/2/82.	325.0	32.0	10.2	450.0	40.1	11.2	350.0	32.0
10/3/82.	144.0	6.9	20.8	481.2	10.8	44.5	396.6	7.8
28/3/82.	80.2	3.2	62.8	145.2	4.4	32.9	101.2	4.4
6/4/82.	22.9	1.6	12.0	40.3	5.4	7.4	20.8	1.4
24/4/82.	5.10	1.0	5.1	7.4	1.6	4.1	4.5	1.2
								3.7

* CF (Concentration factor) = $\frac{\text{concentration in sediment}}{\text{concentration in water}}$

Table 2. Concentration of methyl parathion in fish samples and sea water

Date of Sampling	Fish species	No of Fish	Mean concn. in fish $\mu\text{g} / \text{kg}$	Mean concn. in water	Concentration factor (CF*)
18/2/82.	Sardin	30	131.5	53.9	2.4
	Mugil	30	151.2	53.9	2.8
	Anguilla	30	195.5	53.9	2.6
	Scidena	30	119.0	53.9	2.2
28/2/82.	Sardin	30	149.2	26.7	5.5
	Mugil	30	150.0	26.7	6.0
	Anguilla	30	174.0	26.7	6.5
	Scidena	30	137.8	26.7	5.2
10/3/82.	Sardin	30	49.5	10.47	4.7
	Mugil	30	61.9	10.47	5.9
	Anguilla	30	72.3	10.47	6.9
	Scidena	30	46.7	10.47	4.6

* CF (Concentration Factor) = $\frac{\text{concentration in sediment}}{\text{concentration in water}}$

Neely et al. 1974; Crosby 1975) reported that partition of pesticides between animal fats and the surrounding water depends on the water solubility of the compounds. Such a partition will be independent of fish species in question. Results given by this study, however, showed the variation in the ability of fish species to accumulate the same insecticide even when present in the same aquatic eco-system. Consequently, the metabolic activities and/or excretion may contribute to the final concentration of a given compound in fish species (Eisler 1967).

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